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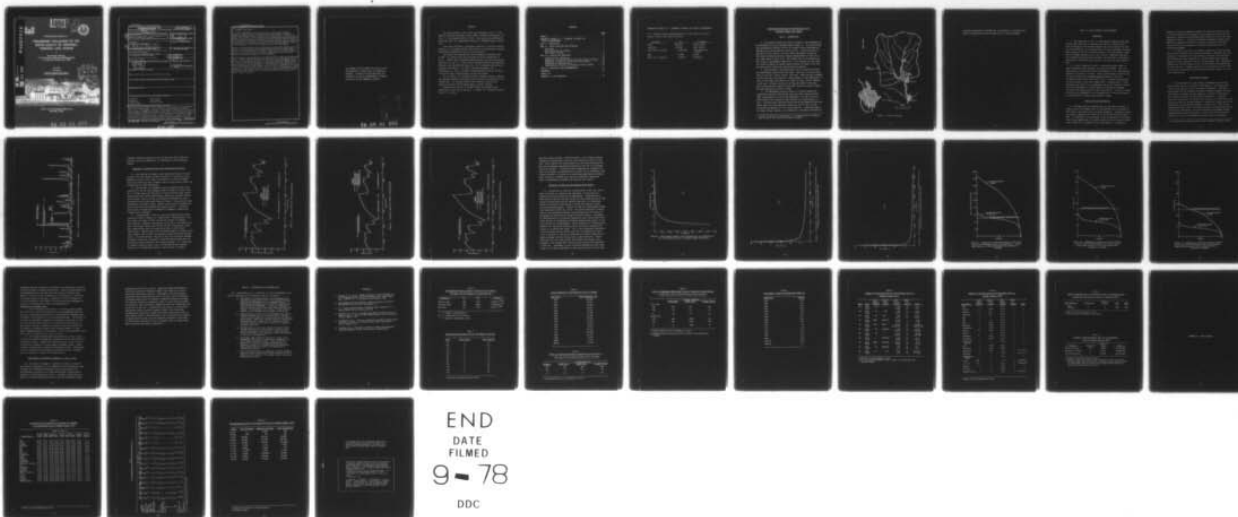
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PRELIMINARY EVALUATION OF THE WATER QUALITY OF PROPOSED TOWANDA LAKE, KANSAS

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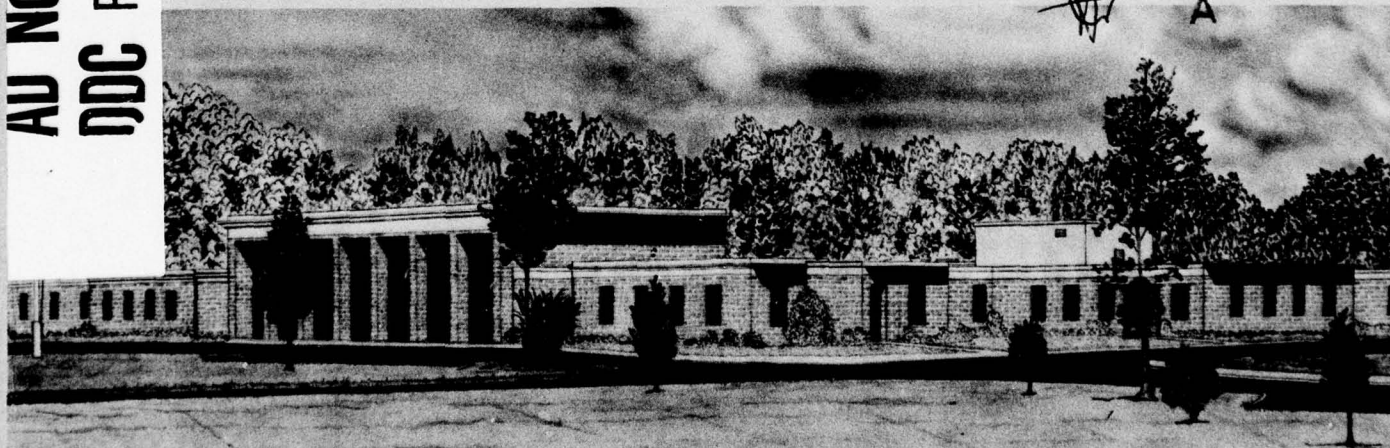
Environmental Laboratory
U. S. Army Engineer Waterways Experiment Station
P. O. Box 631, Vicksburg, Miss. 39180

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| 20. ABSTRACT (Continue on reverse side if necessary and identify by block number) The proposed Towanda Lake would be formed by the impoundment of the Whitewater River near Towanda, Kansas. Project purposes would include flood control, water supply, and general recreation. The purpose of this study was to investigate possible water quality problems that may affect the realization of project benefits—particularly the effects of total dissolved solids (TDS), chloride, and sulfate concentrations on water supply suitability. Because of the high concentrations of these parameters, possible treatment processes were to be evaluated for alleviation of potential water supply problems. (Continued) | | |

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20. ABSTRACT (Continued).

The proposed impoundment would probably be well mixed, exhibiting weak, ephemeral stratification only during periods of calm weather. Therefore, a mass-balance routing of TDS, sulfates, and chlorides through the pool was performed. Daily values of these parameters were generated from continuous specific conductance data and used in conjunction with daily discharge data for input to the routing model. Additionally, data were obtained on trace metals, pesticides, and nutrients for a preliminary evaluation of the possible detrimental effects of these substances on the overall water quality for the impoundment.

Results of the mass-balance routing indicated that only minor problems should be expected with TDS levels and that chloride and sulfate levels should remain below U. S. Environmental Protection Agency (EPA) criteria for these parameters. On the basis of this analysis, no special considerations in water treatment design are considered necessary for these parameters.

Concentrations of mercury and phenol exceeded acceptable levels in the Whitewater River. Although concentrations of these substances in the impoundment are expected to be less than those measured in the river, additional sampling should be conducted to determine the degree of contamination and the expected fate of these constituents in the impoundment. Occasional algal blooms should be expected during periods of calm weather and this should be considered in the design of the water supply withdrawal system. Levels of fecal coliform bacteria in the headwaters of the reservoir should be expected to exceed EPA criteria following storm events, and this may preclude the use of portions of the reservoir for primary contact recreation.

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PREFACE

The work described in this report was performed by the U. S. Army Engineer Waterways Experiment Station (WES), Vicksburg, Mississippi, for the U. S. Army Engineer District, Tulsa. The project was authorized by Intra-Army Order for Reimbursable Services No. TOF770009 dated 15 March 1977.

This report describes a preliminary evaluation of the water quality expected in the proposed Towanda Lake relative to water-quality criteria and standards appropriate for the project purposes.

The project was undertaken by the Environmental Laboratory (EL) at WES. The research was conducted under the direct supervision of Mr. D. L. Robey, Chief, Ecosystem Modeling Branch, and under the general supervision of Drs. R. L. Eley, Chief, Ecosystem Research and Simulation Division, and John Harrison, Chief, EL. Mr. R. Michael Smart served as principal investigator. Dr. Kent W. Thornton and Messrs. N. R. Francingues, Jr., Chief, Treatment Processes Research Branch, and Ross W. Hall provided a detailed review of the draft report.

Mr. Cal Albert, U.S. Geological Survey, Lawrence, Kansas, conducted the sampling program and provided the water-quality data obtained at the Towanda Gage. Mr. Charles Cochran, Tulsa District, conducted the regression analysis and the mass-balance reservoir routing.

Commander and Director of the WES during the preparation and publication of this report was COL J. L. Cannon, CE. Technical Director was Mr. F. R. Brown.

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CONVERSION FACTORS, U. S. CUSTOMARY TO METRIC (SI) UNITS OF MEASUREMENT

U. S. customary units of measurement used in this report can be converted to metric (SI) units as follows:

| <u>Multiply</u> | <u>By</u> | <u>To Obtain</u> |
|-----------------------|------------|----------------------------|
| acres | 4046.856 | square metres |
| acre-feet | 1233.482 | cubic metres |
| cubic feet per second | 0.02831685 | cubic metres per second |
| feet | 0.3048 | metres |
| miles (U. S. statute) | 1.609344 | kilometres |

PRELIMINARY EVALUATION OF THE WATER QUALITY OF
PROPOSED TOWANDA LAKE, KANSAS

PART I: INTRODUCTION

1. The proposed Towanda Lake would be formed by the impoundment of the Whitewater River near Towanda, Kansas (Figure 1). The project would have approximately 200,000 acre-ft* of storage of which approximately 130,000 acre-ft is allocated for flood control. Project purposes, in addition to flood control, are water supply and general recreation. The reservoir would be shallow, having a maximum depth of about 45 ft and a mean depth of about 12 ft. The surface area of the conservation pool would be approximately 6000 acres. It is expected that the reservoir would be similar in thermal characteristics to others in the area and would thus be weakly and intermittently stratified.

2. The watershed is characterized by very low population density and predominantly agricultural land usage. There are only three small towns in the area: Potwin, Benton, and Furley. Wheat and grain sorghum are the principal crops produced. Cattle grazing is also significant in terms of land usage, and there are a number of small, private feedlots in the area. Although drilling in the area was common in the past, these operations have ceased; and the small refinery at Potwin, Kansas, has been inoperable for about a decade.

3. The impoundment would be located in a natural gypsum area resulting in naturally high sulfate (SO_4) and total dissolved solids (TDS) concentrations in runoff and groundwater. The use of brine for extracting oil has resulted in high chloride (Cl) concentrations as well. The Kansas Water Resources Board and the Environmental Protection Agency (EPA) have thus expressed concern over the Cl, SO_4 , and TDS concentrations expected to occur in the reservoir. Since water supply is a major project purpose, it is imperative to determine the degree of

* A table of factors for converting U. S. customary units of measurement to metric (SI) units can be found on page 3.

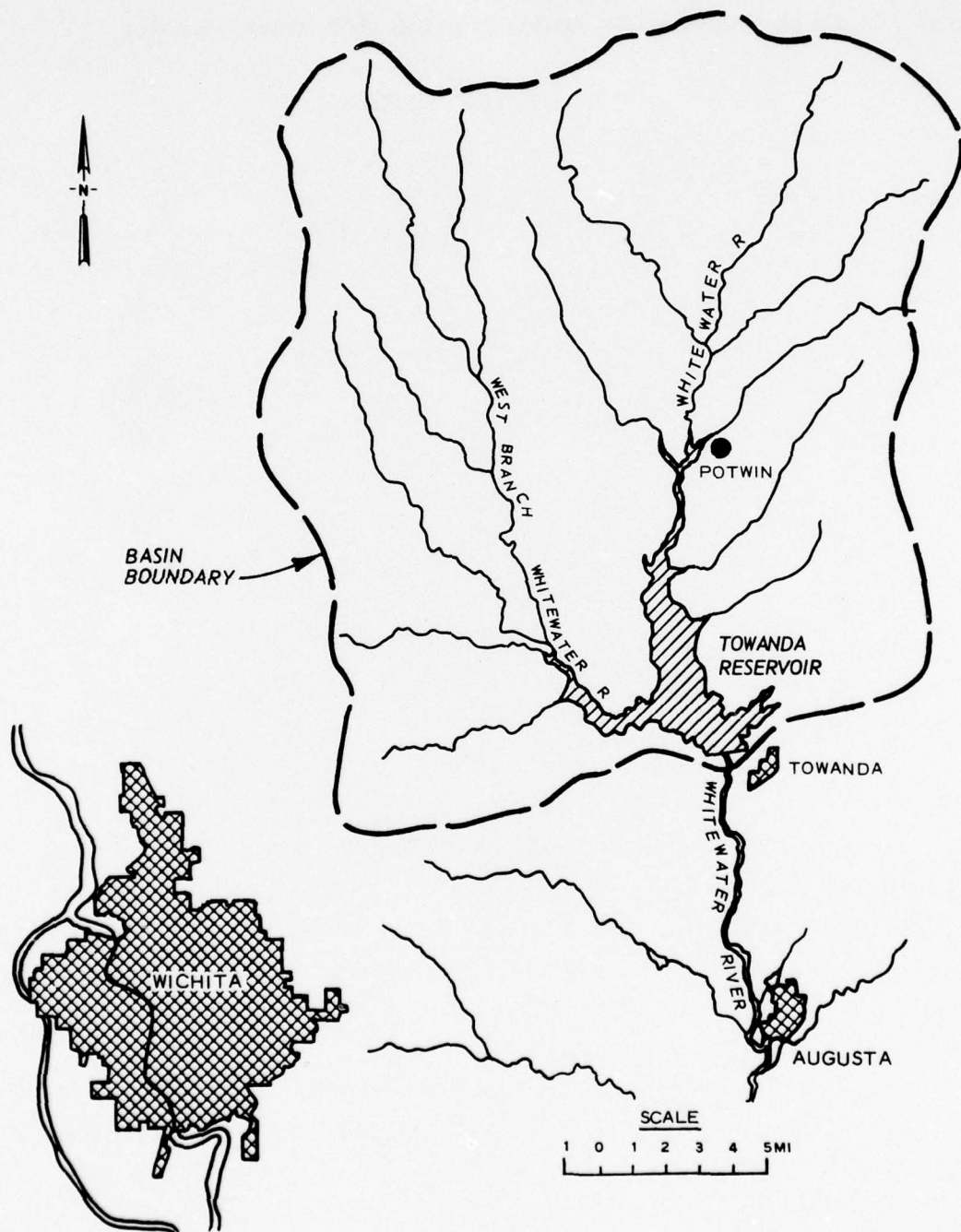


Figure 1. Project area map

potential water-quality problems and, if necessary, to evaluate procedures for dealing with these problems prior to further planning.

PART II: STUDY OBJECTIVES AND PROCEDURES

Objectives

4. The objectives of the U. S. Army Engineer Waterways Experiment Station (WES) study were to determine the extent of expected water-quality problems in the proposed Towanda Lake and to evaluate the need for, and feasibility of, associated water treatment necessary to provide water supply for the city of Wichita, Kansas. The water-quality parameters of primary concern were Cl , SO_4 , and TDS. Additional data were also obtained on heavy metals, nutrients, and pesticides to provide an indication of the overall water quality to be expected in the impoundment.

5. A study conducted by the U. S. Geological Survey (USGS) in 1969¹ indicated serious potential water-quality problems. They concluded that total dissolved solids concentrations in the river exceeded 1000 mg/l 70 percent of the time. River concentrations of sulfates and chlorides exceeded their maximum permissible levels of 250 mg/l for domestic water supply (raw water)² and for finished drinking water³ 79 and 32 percent of the time, respectively. Concentrations of these substances are likely to be considerably less in the reservoir, however, due to the impoundment of large volumes of storm flows containing relatively low concentrations of these constituents. In order to assess the effects of impoundment on the concentrations of these constituents, it was necessary to perform a mass-balance routing.

Data Collection and Analysis

6. Although USGS had determined SO_4 , Cl , and TDS on a monthly to weekly basis at the Towanda Gage, it was desirable for the purposes of this study to have daily values of these parameters. Continuous records of specific conductance were obtained at the Towanda Gage from 1964 to 1969. As conductance is proportional to the concentration of dissolved salts, these numbers give a reliable indication of the relative fluctuations of each of the parameters under investigation. A regression

analysis of specific conductance values with concurrent values for SO_4 , Cl, and TDS indicated a high degree of correlation and resulted in the selection of equations for generating daily values for each of the three parameters (Table 1). Continuous discharge measurements obtained from 1962 to the present by the USGS were used in conjunction with daily values for the above parameters to calculate mass balances for the proposed impoundment.

7. In order to provide a preliminary indication of other possible water-quality problems in the proposed impoundment, additional data were obtained on concentrations of heavy metals, nutrients, and pesticides in the Whitewater River (Appendix A). Monthly samples were obtained during base flow conditions from April through August 1977. Additional samples were obtained during each of two storm events occurring during the study period. A sediment sample was also obtained in May 1977 from a pool at the Towanda Gage and was characterized with respect to pesticide concentrations.

Mass-Balance Routing

8. The assumption that the proposed impoundment would be well mixed allowed the use of a simple routing model for calculating daily values in the lake for each of the chemical parameters. This model incorporated inflows; assumed instantaneous mixing throughout the pool; and allowed for changes in pool volume due to rainfall, evaporation, water-quality releases, and water supply withdrawals. During the initial filling period, no releases occurred until the conservation pool elevation (1285 ft)* had been reached. Conservation storage used in the mass-balance routing was approximately 74,500 acre-ft. When the pool elevation was above 1285 ft, releases were equal to inflows plus precipitation minus evaporation. Below this elevation the minimum release schedule shown in Table 2 was used. In the event the pool elevation decreased to 1275 ft, water-quality releases would be terminated.

* All elevations cited herein are in feet referred to mean sea level.

Conservation storage at this elevation is 28,500 acre-ft. During storm events, releases were not allowed to exceed the downstream channel capacity of 5000 cfs; therefore, the pool elevation increased. This flood storage was released at the maximum amount on subsequent days until the conservation pool elevation was again attained.

PART III: RESULTS AND DISCUSSION

Setting Initial Conditions

9. In order to run the mass-balance routing model, initial concentrations of each of the parameters must be established. As these initial concentrations affect subsequent concentrations, the use of realistic values was imperative. To examine the effect of various initial concentrations, initial filling of the pool was simulated under different conditions. These conditions were selected after examination of annual discharge measurements at the Towanda Gage for the period 1962-1976 (Table 3). Conditions were selected to approximate the range of probable conditions that might occur during filling: a drought period of about 10-year frequency, a median year, and a storm event of 10- to 20-year frequency. The median year was selected on the basis of seasonality of discharge (Figure 2) as well as annual discharge. Water Year (WY) 1967 (the true median) was rejected due to the unusual occurrence of drought conditions early in the year followed by storm events in the latter portion. Water Year 1968 was selected as a median year due to the occurrence of normal seasonal peaks in discharge.

10. Filling was accomplished by impounding all inflowing water until the conservation pool elevation (1285 ft) was attained. This process took approximately 1 year for the median year and 6 days for the storm event filling. The conservation pool elevation was never attained during the drought period due to insufficient inflow. Concentrations of TDS, SO_4 , and Cl achieved in this manner are representative of the maximum range of initial conditions expected in the impoundment (Table 4).

11. Table 5 shows the convergence of concentrations for each of the parameters during two subsequent median-year cycles. After two median-year cycles, concentrations were not appreciably different regardless of filling conditions. Therefore, concentrations attained by filling during a median year followed by two median-year cycles were selected as initial conditions. These concentrations represent

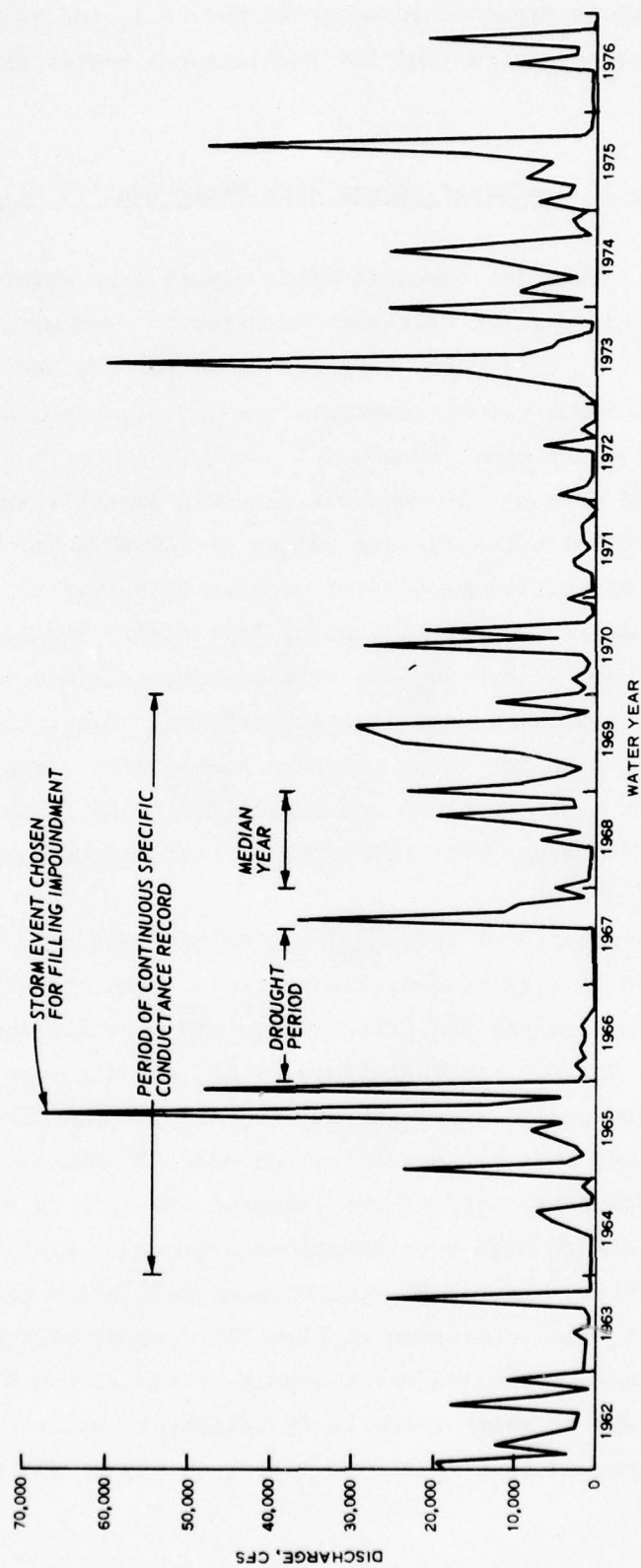


Figure 2. Discharge record for the Whitewater River at Towanda, Kansas

"average" conditions expected to occur in the pool, and as such, are realistic initial concentrations for simulating a 6-year hydrologic routing.

Comparison of Predicted Values with Water-Quality Criteria

12. EPA criteria for domestic water supply (raw water) are incorporated in the EPA "Quality Criteria for Water."² Criteria for those substances listed in this report are identical for raw and finished waters as conventional water treatment does not significantly affect the concentration of these constituents.

13. The EPA does not recommend a specific domestic water supply criterion for TDS but cites maximum values of 250 mg/l for both Cl and SO₄.² Problems associated with water of high dissolved solids content include unpleasant tastes, physiological (especially laxative) effects, and corrosion. Both Cl and SO₄ are primarily responsible for taste problems while SO₄ alone causes laxative effects. Concentrations below the 200-mg/l criterion for these elements should afford reasonable protection from both taste problems and laxative effects.² The Public Health Service "Drinking Water Standards"³ recommends a maximum TDS concentration of 500 mg/l.

14. Concentrations of TDS, SO₄, and Cl obtained from the 6-year routing are shown in Figures 3-5, respectively. Concentrations of SO₄ were always well below the EPA criterion of 250 mg/l for domestic water supplies. Maximum concentrations of SO₄ and Cl, occurring at the end of the drought period, were 190 and 125 mg/l, respectively, while mean concentrations over the 6-year period were 131 and 74 mg/l, respectively. Concentrations of TDS exceeded 500 mg/l 37 percent of the time, primarily due to high concentrations achieved during the drought period. As concentrations of SO₄ and Cl were well below their maximum acceptable values, the occurrence of high TDS concentrations does not necessarily indicate potential water supply problems. In a study based on consumer surveys of water taste in 29 California water systems, Bruvold et al.⁴ showed that waters of less than 659 to 755 mg/l

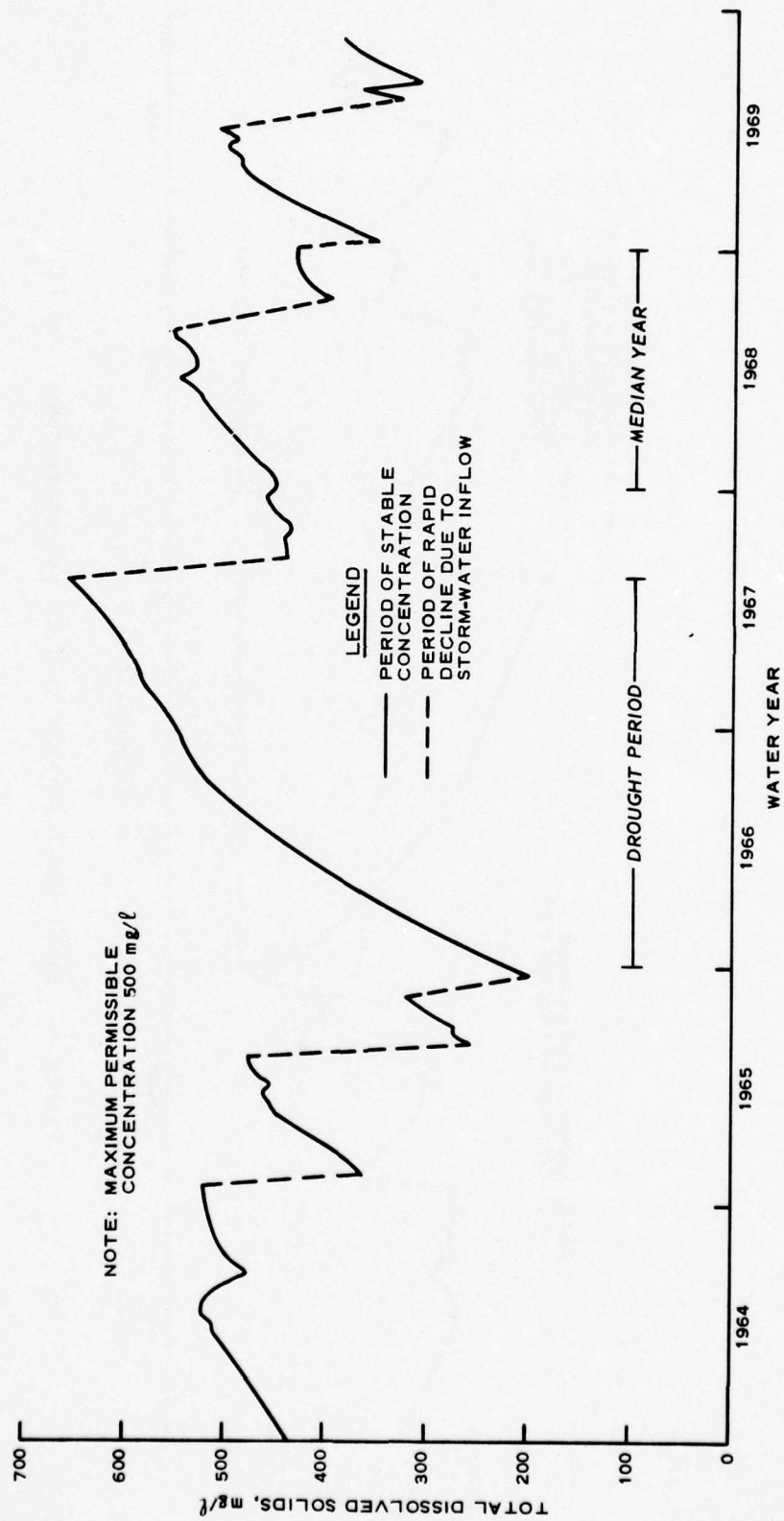


Figure 3. Predicted reservoir TDS concentrations

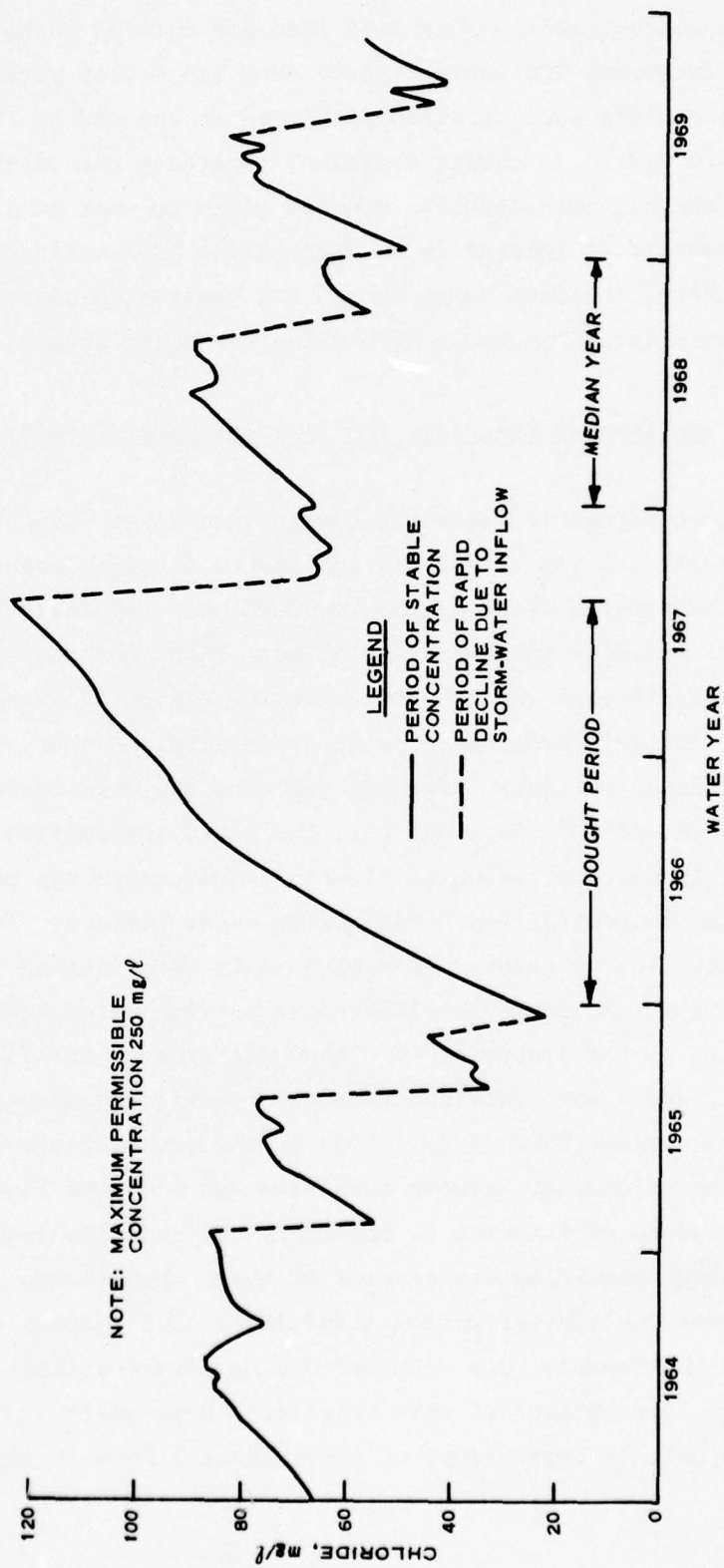


Figure 5. Predicted reservoir chloride concentrations

dissolved solids received a rating of "good." In the 6-year reservoir routing, TDS concentrations were less than 658 mg/l 99 percent of the time. Mean reservoir TDS concentration over the 6-year period was 462 mg/l and the maximum concentration predicted at the end of the drought period was 660 mg/l. In making decisions regarding the suitability of water for drinking, consideration must be given to user acclimation. The proposed reservoir is located in an area having high-salinity drinking water; therefore, resident users should not experience significant taste or physiological problems from using the water supply.

Comparison of Reservoir and Riverine Water Quality

15. Fluctuations in reservoir concentrations of TDS, SO_4 , and Cl shown in Figures 3-5 illustrate the importance of storm events in regulating the water quality of the impoundment. Concentrations of these species steadily increased during base flow conditions due to high influent concentrations and the concentrating effect of evaporation. During the relatively brief periods of storm flow, however, concentrations decreased markedly. Figures 6-8 show the relationships between river discharge and TDS, SO_4 , and Cl. The rapid attenuation of concentration with increasing discharge clearly demonstrates the reason for declining pool concentrations during storm-water inflows. Impoundment of large quantities of relatively high-quality water during these storm events results in considerable differences between water quality of the river and that of the impoundment. These differences are illustrated in Figures 9-11, where data obtained from the routing are compared with results of an earlier USGS study.¹ Due to the considerable differences in riverine water quality between base flow and elevated flow conditions, it would be of interest to determine the relative quantities of annual discharge occurring during each of these conditions. Table 6 shows that over the 15-year period (1962-1976) 55.3 percent of the discharge at the Towanda Gage occurred during elevated flow conditions (> 2000 cfs). Impoundment of this relatively high-quality water would result in an overall improvement of water quality both in the pool and

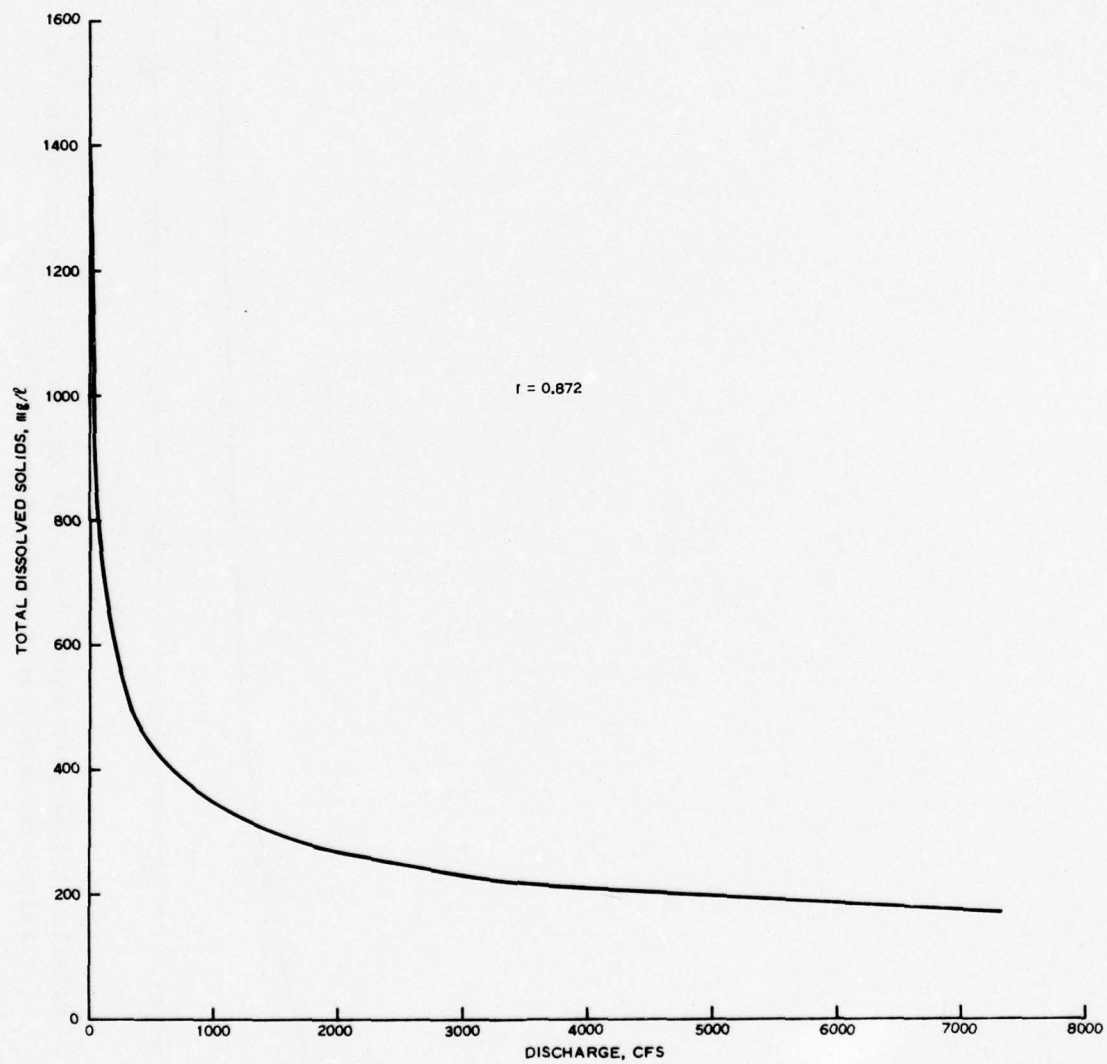


Figure 6. Relationship between river discharge and TDS concentration for the Whitewater River at Towanda, Kansas (1962-1975)

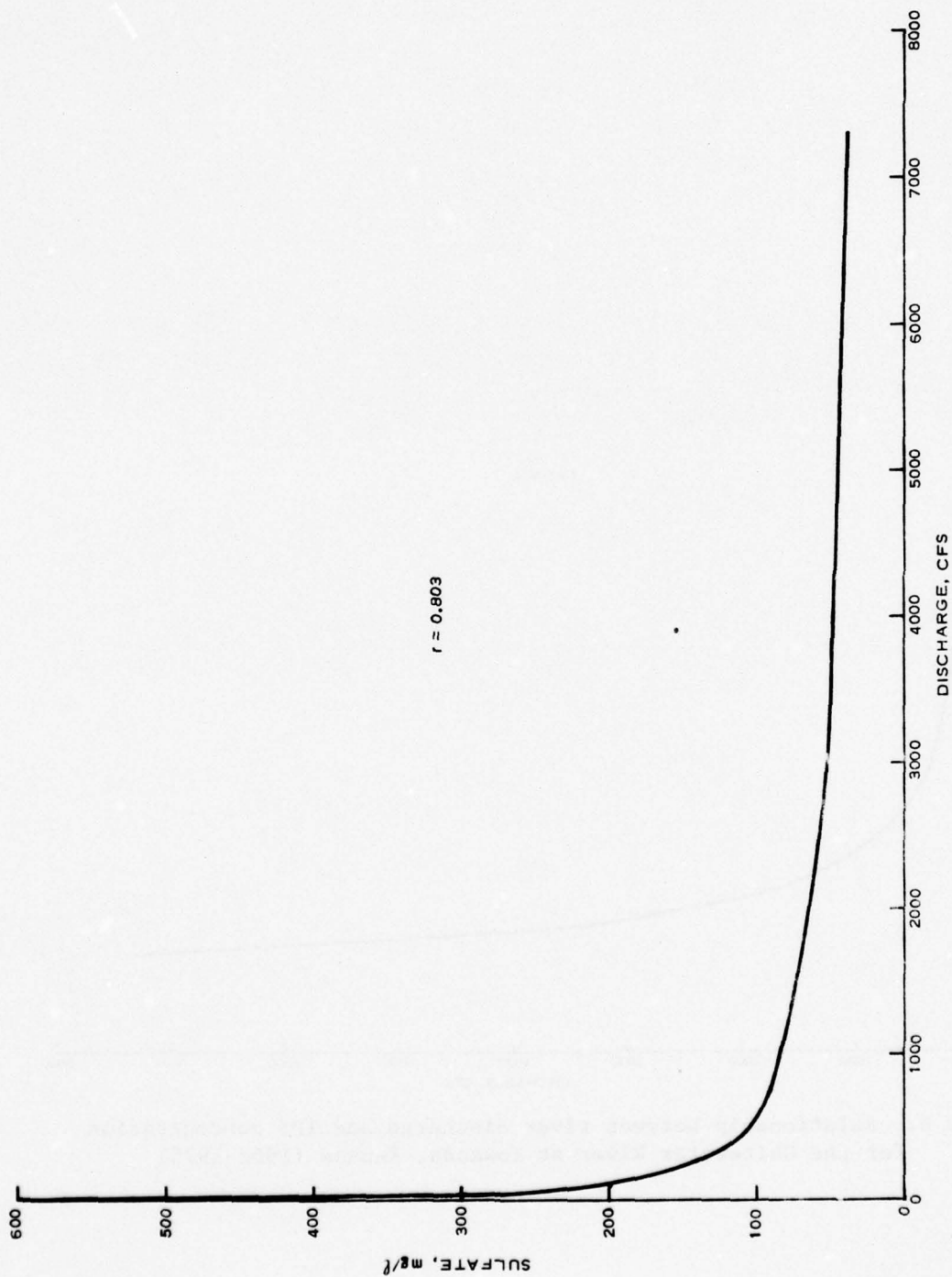


Figure 7. Relationship between river discharge and sulfate concentration for the Whitewater River at Towanda, Kansas (1962-1975)

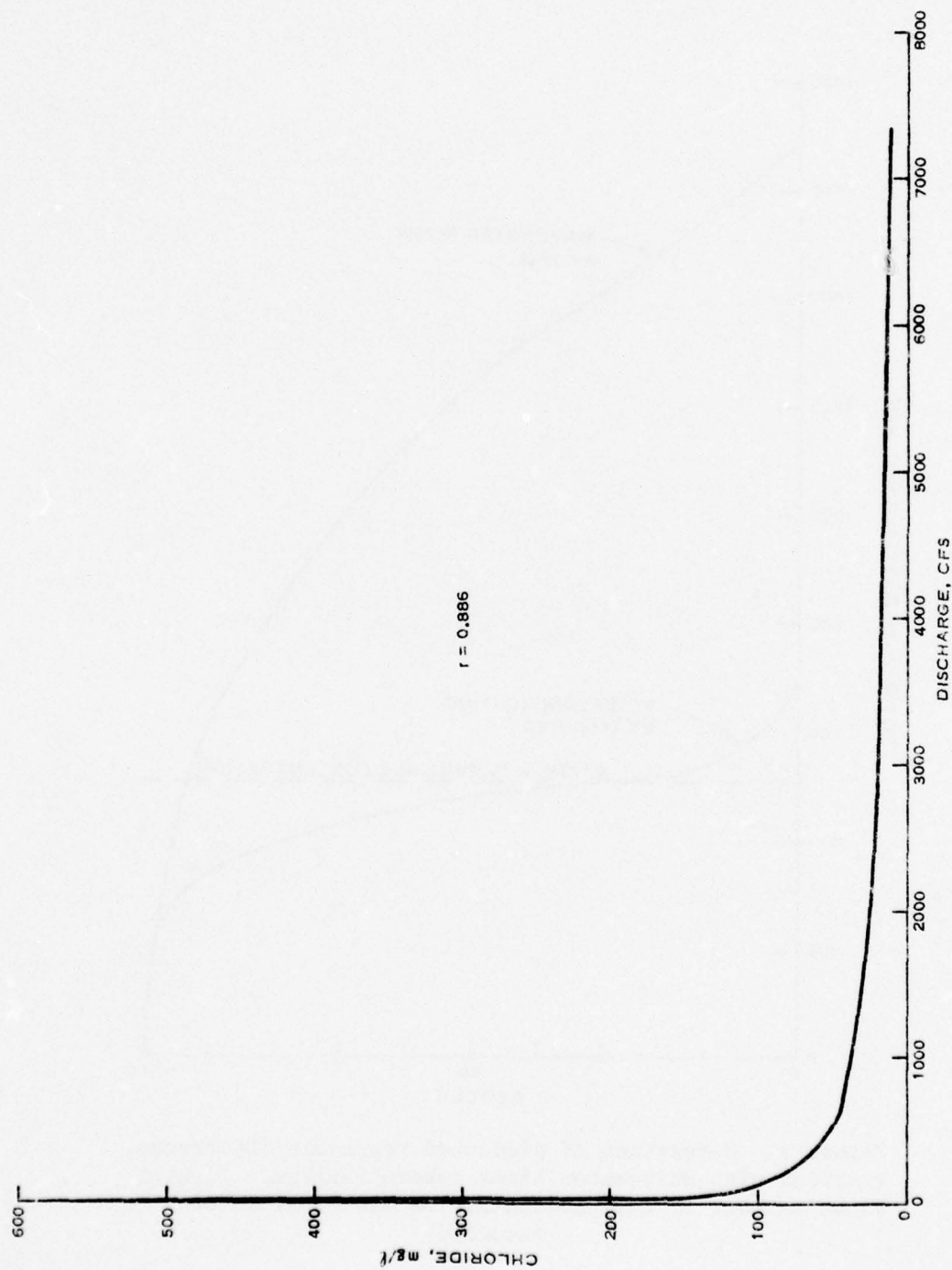


Figure 8. Relationship between river discharge and chloride concentration for the Whitewater River at Towanda, Kansas (1962-1975)

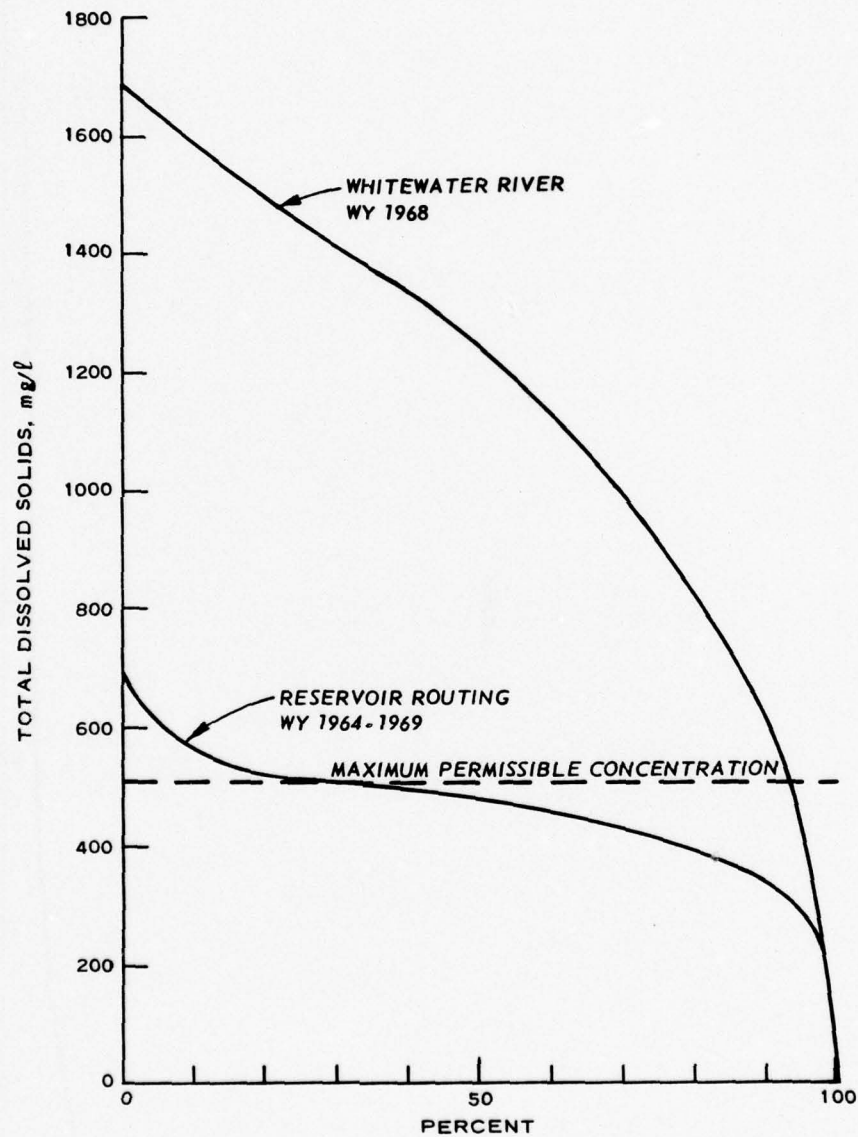


Figure 9. Comparison of predicted reservoir TDS concentrations with Whitewater River concentrations. Figure shows percent of time concentration was equalled or exceeded

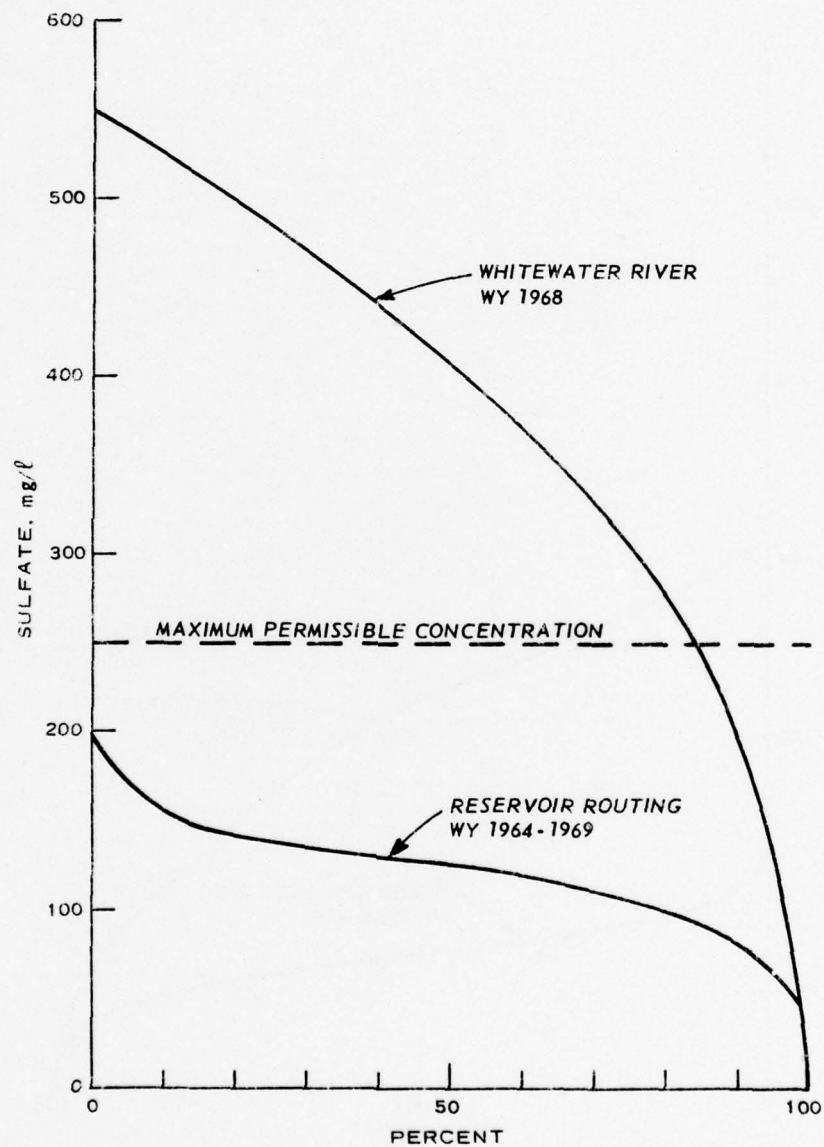


Figure 10. Comparison of predicted reservoir sulfate concentrations with Whitewater River concentrations. Figure shows percent of time concentration was equalled or exceeded

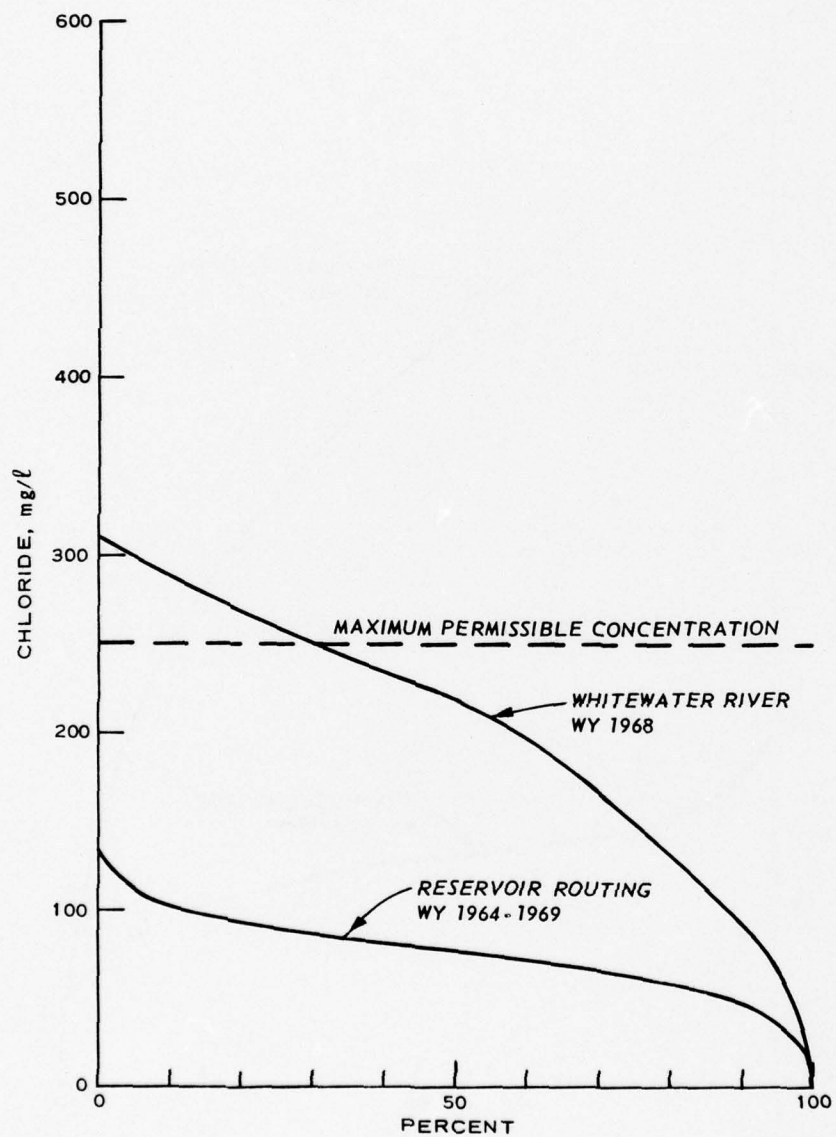


Figure 11. Comparison of predicted reservoir chloride concentrations with Whitewater River concentrations. Figure shows percent of time concentration was equalled or exceeded

downstream from the dam and would provide a buffer against the poor water quality occurring during base flow conditions.

Evaluation of Additional Data

Trace metals

16. Trace metal concentrations in the Whitewater River at Towanda are summarized in Table 7 along with the appropriate EPA criteria. Complete data are presented in Tables A1 and A2. Of the trace metals analyzed in the river, only cadmium and manganese occasionally exceeded EPA criteria for public water supply. Mean river concentrations of dissolved cadmium were 2 $\mu\text{g}/\ell$ during base flow and 7 $\mu\text{g}/\ell$ during elevated flow conditions. Mean dissolved manganese concentrations were 178 and 24 $\mu\text{g}/\ell$ during base and elevated flows, respectively. Concentrations of both of these constituents are expected to be lower in the reservoir than in the river.

17. Both total and dissolved forms of mercury frequently exceeded the EPA criteria of 0.05 $\mu\text{g}/\ell$ for the protection of aquatic life, but were less than the criteria of 2.0 $\mu\text{g}/\ell$ for domestic water supply.² The occurrence of mercury under both base and elevated flow conditions indicates a probable loading to the impoundment in excess of allowable limits for protection of aquatic life. Additional data should be obtained on mercury concentrations in the river and it would also be desirable to determine mercury concentrations in tissues of organisms inhabiting the river to ascertain the degree of accumulation within different trophic levels. If body burden concentrations in reservoir aquatic life exceed Food and Drug Administration limits, some aspects of the recreational potential of the project would be restricted.

Pesticides

18. Analysis of pesticide concentrations (Table 8) revealed concentrations below detection limits for all but 2,4-D and 2,4,5-T. These pesticides are not expected to cause problems in the proposed impoundment either for water supply or recreation purposes. River

concentrations of phenols, however, were consistently above the EPA water supply criterion of 1 µg/l (Table 9). Phenolic compounds arise from oil refineries, livestock dips, degradation of pesticides, and natural sources. While phenol is rapidly degraded in natural water, some of these compounds, depending on the nature of the substituents on the benzene ring, are highly refractory and may exist for extended periods prior to degradation. Phenols are not removed efficiently by conventional water treatment and can be chlorinated during treatment to produce malodorous compounds.² Threshold odor levels can be as low as 2 µg/l for 2-Chlorophenol.⁵ Phenolic compounds can also cause tainting of fish flesh. Additional data should be obtained on phenol concentrations in the Whitewater River and in fish inhabiting the river to ascertain the degree of accumulation.

Nutrients and eutrophication

19. While an evaluation of the potential for algal blooms in the proposed reservoir is beyond the scope of this study, measured river concentrations of nitrogen and phosphorus (Table A4) are sufficient to support nuisance algal blooms. As the pool is not expected to exhibit strong stratification, wind action should maintain high turbidity. Algal populations are therefore expected to be light limited and blooms should occur only during periods of calm weather and lower turbidity. The occurrence of algal blooms is expected to be similar in frequency, magnitude, and duration to those of other shallow impoundments in the surrounding area.

Coliform bacteria

20. Fecal coliforms measured in the Whitewater River (Table 10) greatly exceeded the EPA criteria for primary contact recreation.² The EPA does not give a criterion for domestic water supply as coliform bacteria are easily controllable with conventional water supply treatment (chlorination). It is beyond the scope of this study to evaluate the extent of this problem in the reservoir, but the possibility exists that portions of the reservoir may not be useable for primary contact recreation following storm events. High levels of coliform bacteria will enter the reservoir during storm flows and will be distributed

throughout the pool according to the quantity and duration of the storm flow and the degree of mixing and dilution. Levels of these bacteria in the headwaters of the pool may exceed EPA criteria following storm events. The source of these bacteria will be difficult to control due to the diffuse nature of the input. Comparison of fecal coliforms (FC) to fecal streptococci (FS) ($FC/FS = 0.5$) indicates the predominance of livestock or poultry wastes rather than human wastes.⁶

Estimation of treatment needs

21. Since predicted concentrations of Cl and SO_4 never exceeded EPA criteria, no specialized treatment for these parameters should be required. Predicted concentrations of TDS exceeded the 500-mg/l level 37 percent of the time but in no case exceeded 660 mg/l. The quality of the water for drinking is expected to be marginal with respect to TDS, but the expected concentrations do not justify the added expense of specialized treatment. Possible treatment processes for reducing TDS that might be considered during further planning, however, include partial demineralization or distillation. Both of these processes would also result in reduction of Cl , SO_4 , Ca , and Mg .

22. Water supply problems associated with trace metals or pesticides are not expected although Mn concentrations in the hypolimnion may exceed EPA criteria. Consideration should be given to this potential problem in the design of the water supply withdrawal system. Concentrations of phenols may occasionally exceed EPA criteria. Chlorination of these compounds during treatment may produce malodorous compounds that are difficult to remove from finished waters.

Applicability of Results to Changes in Project Design

23. The effect of changes in conservation storage or yield on water-quality constituents was not addressed in this preliminary study. Increases in conservation storage may increase the likelihood of stratification, in which case a simple mass-balance routing would be inadequate for assessing water quality. Increased storage would also result in increased evaporation from the pool and consequently higher

concentrations of TDS, SO_4 , and Cl. Under the conditions examined in this study, inflows are frequently smaller than losses due to evaporation and water-quality releases. Consequently, the pool elevation continually drops during drier periods of the year, and the occurrence of seasonal storm events is necessary to maintain conservation storage. The resultant increase in evaporative losses of water from an increase in storage may cause considerable changes in water level between dry and wet periods of the year. Increasing the water supply or water-quality releases may result in similar changes in water level fluctuation. Decreasing the conservation storage would lessen evaporative losses of water and may result in decreased water level fluctuations and, possibly, in improved water quality over the conditions evaluated in this study. The exact nature and extent of any such changes could be readily determined using the mass-balance routing model.

PART IV: CONCLUSIONS AND RECOMMENDATIONS

24. Conclusions and recommendations concerning impoundment water quality based upon this preliminary study are as follows:

- a. The results of the mass-balance routing indicate that expected concentrations of TDS in the proposed impoundment would occasionally exceed recommended limits. No problems are anticipated with Cl and SO₄ concentrations in the proposed impoundment. Analysis of additional data did not indicate serious water supply problems with respect to trace metals or pesticides. Based on these results, no specialized water treatment is warranted.
- b. Concentrations of mercury in the impoundment may exceed levels considered safe for the protection of aquatic life but are not expected to exceed criteria for domestic water supply. Additional sampling should be conducted to determine mercury loadings to the impoundment as well as the accumulation of mercury in tissues of organisms inhabiting the river.
- c. Concentrations of phenol may exceed acceptable levels and may cause tainting of fish flesh as well as taste and odor problems in finished water. Additional data should be obtained on the loading of phenol and possibly on the rate of decomposition.
- d. Occasional algal blooms are expected to occur in the impoundment and should be considered in designing the water supply withdrawal system. Algal blooms are expected to be similar in frequency, magnitude, and duration to those of surrounding lakes and reservoirs. Algal blooms should not affect the realization of recreational benefits.
- e. Fecal coliform bacteria are expected to exceed allowable levels following storm events and may preclude the use of portions of the reservoir for primary contact recreation during these periods.

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6. Millipore Corp., "Biological Analysis of Water and Wastewater," Application Manual AM 302, 1973, Bedford, Massachusetts.

Table 1
Relationships Between Specific Conductance and Sulfates,
Chlorides, and TDS During the Period 1962-1974

| <u>Parameter</u> | <u>n*</u> | <u>r**</u> | <u>Function</u> |
|------------------|-----------|------------|---------------------|
| Sulfate, mg/l | 300 | 0.947 | $0.0546 C^{1.168+}$ |
| Chloride, mg/l | 402 | 0.982 | $0.00458 C^{1.447}$ |
| TDS, mg/l | 253 | 0.993 | $0.811 C^{0.968}$ |

* n = number of observations.

** r = coefficient of correlation.

+ C = specific conductance, μ mhos.

Table 2
Minimum Release Schedule Used in Mass-Balance Routing*

| <u>Month</u> | <u>Water Supply</u> | <u>Water Quality</u> |
|--------------|---------------------|----------------------|
| Oct | 3 | 22 |
| Nov | 3 | 19 |
| Dec | 3 | 19 |
| Jan | 3 | 19 |
| Feb | 3 | 15 |
| Mar | 3 | 19 |
| Apr | 3 | 22 |
| May | 3 | 22 |
| Jun | 3 | 28 |
| Jul | 3 | 35 |
| Aug | 3 | 32 |
| Sep | 3 | 24 |

* Units are in cubic feet per second.

Table 3
Annual Discharge of the Whitewater River at Towanda

| <u>Water Year</u> | <u>Total Discharge, cfs</u> |
|-------------------|-----------------------------|
| 1962 | 104,596 |
| 1963 | 38,713 |
| 1964 | 20,681 |
| 1965 | 181,618 |
| 1966 | 15,002 |
| 1967 | 61,038 |
| 1968 | 52,456 |
| 1969 | 165,194 |
| 1970 | 67,423 |
| 1971 | 16,975 |
| 1972 | 18,374 |
| 1973 | 124,521 |
| 1974 | 113,878 |
| 1975 | 103,987 |
| 1976 | 41,150 |
| MEAN | 75,040 |
| MEDIAN | 61,038 |

Table 4
Effect of Filling Conditions on Reservoir Concentrations of
TDS, SO₄, and Cl Achieved During Initial Filling*

| <u>Parameter</u> | <u>Storm Event</u> | <u>Filling Condition</u> | |
|------------------|--------------------|--------------------------|-----------------------|
| | | <u>Median Year</u> | <u>Drought Period</u> |
| TDS | 125 | 464 | 1485 |
| SO ₄ | 27 | 125 | 453 |
| Cl | 10 | 74 | 315 |

* Concentrations are in micrograms per litre.

Table 5
Effect of Subsequent Median-Year Cycles on Reservoir Concentrations
of TDS, SO₄, and Cl Achieved During Initial Filling*

| | <u>Filling Condition</u> | | |
|--------------------|--------------------------|--------------------|-----------------------|
| | <u>Storm Event</u> | <u>Median Year</u> | <u>Drought Period</u> |
| <i>First Year</i> | | | |
| TDS | 347 | 437 | 623 |
| SO ₄ | 89 | 115 | 175 |
| Cl | 50 | 67 | 111 |
| <i>Second Year</i> | | | |
| TDS | 406 | 430** | 479 |
| SO ₄ | 105 | 112** | 128 |
| Cl | 61 | 65** | 77 |

* Concentrations are in micrograms per litre.

** Concentrations selected as initial conditions for mass-balance routing.

Table 6

Percentage of Annual Discharge Above 2000 cfs

| <u>Water Year</u> | <u>Percent</u> |
|-------------------|----------------|
| 1962 | 46.4 |
| 1963 | 50.4 |
| 1964 | 16.4 |
| 1965 | 69.8 |
| 1966 | 0.0 |
| 1967 | 69.7 |
| 1968 | 44.4 |
| 1969 | 59.9 |
| 1970 | 48.0 |
| 1971 | 0.0 |
| 1972 | 0.0 |
| 1973 | 65.0 |
| 1974 | 54.5 |
| 1975 | 53.9 |
| 1976 | 68.3 |
| 1964-69 | 59.5 |
| 1962-76 | 55.3 |

Table 7
Summary of Trace Metal Data for the Whitewater River at
Towanda, Kansas, 1977*

| <u>Metal</u> | <u>Phase</u> | <u>Water Supply Criteria</u> | <u>Aquatic Life Criteria</u> | <u>Mean Concen- tration</u> | <u>Number of Samples</u> | <u>Range</u> |
|--------------|--------------|--------------------------------------|--------------------------------------|-------------------------------------|----------------------------------|--------------|
| As | Diss. | 50 | -- | 2.60 | 13 | 2-7 |
| | Total | | | 10.68 | 13 | <1-34 |
| Cd | Diss. | 10 | 12** | 5.80 | 13 | 1-13 |
| | Total | | | 10.00 | 13 | <10-10 |
| Cr | Diss. | 50 | 100 | 1.45 | 13 | <1-4 |
| | Total | | | 18.77 | 13 | <10-40 |
| Cu | Diss. | 1000 | Bioassay | 8.20 | 1 | 2-19 |
| | Total | | | 21.60 | 1 | 1-40 |
| Fe | Diss. | 300 | 1000 | 125.40 | 13 | 20-280 |
| | Total | | | 14,923.00 | 13 | 1300-36,000 |
| Pb | Diss. | 50 | Bioassay | 7.00 | 13 | 1-22 |
| | Total | | | 100.00 | 13 | <100-100 |
| Mn | Diss. | 50 | -- | 65.10 | 13 | <1-470 |
| | Total | | | 676.90 | 13 | 210-1500 |
| Ni | Diss. | -- | Bioassay | 4.00 | 13 | 2-7 |
| | Total | | | 50.00 | 13 | <50-50 |
| Zn | Diss. | 5000 | Bioassay | 20.20 | 13 | 4-50 |
| | Total | | | 76.90 | 13 | 30-160 |
| Se | Diss. | 10 | Bioassay | 1.08 | 13 | <1-2 |
| | Total | | | 1.41 | 13 | <1-4 |
| Hg | Diss. | 2.0 | 0.05 | 0.23 | 13 | <0.1-0.6 |
| | Total | | | 0.20 | 12 | <0.1-0.6 |

* Units are in micrograms per litre.

** For less sensitive aquatic life; 1.2 µg/l for cladocerans and salmonid fishes.

Table 8
Summary of Pesticide Data for the Whitewater River at
Towanda, Kansas, 1977*

| <u>Constituent</u> | <u>Water Supply Criteria</u> | <u>Aquatic Life Criteria</u> | <u>Mean Concen- tration</u> | <u>Number of Samples</u> | <u>Range</u> |
|------------------------|--------------------------------------|--------------------------------------|-------------------------------------|----------------------------------|--------------|
| PCN | -- | -- | <0.01 | 7 | -- |
| Aldrin | -- | 0.003 | <0.01 | 7 | -- |
| Lindane | 4.0 | 0.01 | <0.01 | 7 | -- |
| Chlordane | -- | 0.01 | <0.10 | 7 | -- |
| DDD | -- | -- | <0.01 | 7 | -- |
| DDE | -- | -- | <0.01 | 7 | -- |
| DDT | -- | 0.001 | <0.01 | 7 | -- |
| Dieldrin | -- | 0.003 | <0.01 | 7 | -- |
| Endosulfan | -- | 0.003 | <0.01 | 7 | -- |
| Endrin | 0.2 | 0.004 | <0.01 | 7 | -- |
| Ethion | -- | -- | <0.01 | 7 | -- |
| Toxaphene | 5 | 0.005 | <0.10 | 7 | -- |
| Heptachlor | -- | 0.001 | <0.01 | 7 | -- |
| Heptachlor- Epoxide | -- | -- | <0.01 | 7 | -- |
| PCB | -- | 0.001 | <0.10 | 7 | -- |
| Malathion | -- | 0.01 | <0.01 | 7 | -- |
| Parathion | -- | 0.04 | 0.01 | 7 | <0.01-0.01 |
| Diazinon | -- | -- | 0.03 | 7 | <0.01-0.12 |
| Methylpara- thion | -- | -- | <0.01 | 7 | -- |
| 2,4-D | 100 | -- | 0.27 | 6 | 0.08-0.45 |
| 2,4,5-T | 10 | -- | 0.17 | 6 | 0.03-0.49 |
| Silvex | -- | -- | <0.01 | 6 | -- |
| Trithion | -- | -- | <0.01 | 7 | -- |
| Methyltrithion | -- | -- | 0.01 | 7 | <0.01-0.01 |

* Units are in micrograms per litre.

Table 9
Phenol Concentrations in the Whitewater River at Towanda, Kansas,
During Base Flow and Elevated Flow Conditions, 1977*

| <u>Flow Condition</u> | <u>Criterion**</u> | <u>Number of Samples</u> | <u>Mean</u> | <u>Range</u> |
|-----------------------|--------------------|------------------------------|-------------|--------------|
| Base | 1 | 4 | 3.25 | 1-7 |
| Elevated | 1 | 9 | 4.44 | 2-14 |

* Units are in micrograms per litre.

** EPA criterion for domestic water supply.

Table 10
Summary of Microbiological Data for the Whitewater
River at Towanda, Kansas, 1977*

| <u>Parameter</u> | <u>Criteria**</u> | <u>Geometric Mean</u> | <u>Range</u> |
|---------------------|-------------------|---------------------------|--------------|
| Fecal coliforms | 200 | 9,800 | 60-110,000 |
| Immediate coliforms | -- | 41,500 | 750-460,000 |
| Fecal streptococci | -- | 18,100 | 220-370,000 |

* Values are colonies per millilitres.

** The EPA criteria for primary contact recreation state that fecal coliform levels should not exceed a geometric mean of 200 per 100 ml nor should more than 10 percent of the samples taken during any 30-day period exceed 400 per 100 ml.

APPENDIX A: DATA COLLECTED

Table A1
Trace Metal Concentrations in the Whitewater River at Towanda, Kansas,
During Base Flow Conditions*

| <u>Trace Metal</u> | <u>Phase</u> | <u>Date</u> | | | | |
|--------------------|--------------|---------------|---------------|---------------|---------------|--------------|
| | | <u>21 Apr</u> | <u>19 May</u> | <u>27 Jun</u> | <u>21 Jul</u> | <u>4 Aug</u> |
| As | Total | 6 | -- | 4 | 2 | 6 |
| | Diss. | 5 | -- | 4 | 2 | 2 |
| Cd | Total | 10 | <10 | 10 | <10 | <10 |
| | Diss. | 2 | 1 | 3 | 2 | 2 |
| Cr | Total | <10 | <10 | 10 | 10 | 10 |
| | Diss. | < 1 | < 1 | < 1 | < 1 | < 1 |
| Cu | Total | 20 | 10 | <10 | 10 | 1 |
| | Diss. | 5 | 1 | 3 | 2 | 2 |
| Fe | Total | 1300 | 1600 | 3200 | 1300 | 4600 |
| | Diss. | 140 | 20 | 30 | 20 | 30 |
| Pb | Total | 100 | <100 | 100 | <100 | <100 |
| | Diss. | 4 | 8 | 5 | 9 | 5 |
| Mn | Total | 630 | 690 | 210 | 220 | 300 |
| | Diss. | 470 | 260 | 50 | 10 | 100 |
| Ni | Total | <50 | <50 | <50 | <50 | <50 |
| | Diss. | 2 | 4 | 3 | 2 | 3 |
| Zn | Total | 30 | 30 | 70 | 50 | 30 |
| | Diss. | 10 | 20 | 10 | 4 | 8 |
| Se | Total | 1 | 1 | 2 | 1 | 2 |
| | Diss. | 1 | 1 | 2 | 1 | 1 |
| Hg | Total | <0.1 | 0.1 | <0.1 | -- | 0.1 |
| | Diss. | <0.1 | 0.3 | <0.1 | 0.4 | 0.2 |

* Units are in micrograms per litre.

Table A2
Trace Metal Concentrations in the Whitewater River at Towanda, Kansas,
Under Elevated Flow Conditions, 1977*

| Trace Metal | Phase | Date and Time, hr | | | | | | | | | | | |
|-------------|-------|-------------------|-------|-------|--------|-------|-------|--------|------|------|--------|------|------|
| | | 21 May | | | 22 May | | | 11 Aug | | | 12 Aug | | |
| | | 1500 | 1900 | 0855 | 1325 | 1700 | 1325 | 1610 | 2015 | 0820 | 1010 | 1230 | 1410 |
| As | Total | 9 | 6 | 13 | 22 | 3 | 34 | -- | -- | 15 | -- | <1 | 18 |
| | Diss. | 2 | 2 | 3 | 2 | 2 | 2 | -- | -- | 4 | -- | 2 | 2 |
| Cd | Total | <10 | <10 | <10 | <10 | <10 | <10 | -- | -- | <10 | -- | <10 | <10 |
| | Diss. | 13 | 9 | 13 | 5 | 2 | 8 | -- | -- | 4 | -- | 7 | 5 |
| Cr | Total | 20 | 40 | 20 | 30 | 0 | 36 | -- | -- | 16 | -- | 16 | 16 |
| | Diss. | <1 | <1 | <1 | <1 | <1 | 4 | -- | -- | 4 | -- | <1 | <1 |
| Cu | Total | 30 | 40 | 30 | 30 | 20 | 40 | -- | -- | 20 | -- | 10 | 20 |
| | Diss. | 5 | 11 | 15 | 5 | 5 | 12 | -- | -- | 11 | -- | 19 | 12 |
| Fe | Total | 23000 | 36000 | 29000 | 25000 | 21000 | 24000 | -- | -- | 9200 | -- | 7200 | 9200 |
| | Diss. | 280 | 110 | 160 | 130 | 160 | 150 | -- | -- | 160 | -- | 120 | 140 |
| Pb | Total | <100 | 100 | <100 | 100 | 100 | <100 | -- | -- | <100 | -- | <100 | <100 |
| | Diss. | 9 | 4 | 6 | 2 | 1 | 11 | -- | -- | 2 | -- | 11 | 22 |
| Mn | Total | 790 | 1400 | 850 | 690 | 610 | 1500 | -- | -- | 530 | -- | 530 | 540 |
| | Diss. | 120 | 40 | 8 | <1 | 8 | 20 | -- | -- | 20 | -- | <1 | <1 |
| Ni | Total | <50 | <50 | <50 | 50 | 50 | 50 | -- | -- | <50 | -- | <50 | 50 |
| | Diss. | 6 | 6 | 7 | 4 | 2 | 5 | -- | -- | 4 | -- | 4 | 4 |
| Zn | Total | 90 | 160 | 100 | 110 | 70 | 130 | -- | -- | 50 | -- | 50 | 60 |
| | Diss. | 20 | 20 | 20 | 30 | 50 | 20 | -- | -- | 20 | -- | 30 | 20 |
| Se | Total | 1 | 1 | 1 | <1 | <1 | <1 | -- | -- | 4 | -- | 1 | <1 |
| | Diss. | <1 | <1 | <1 | <1 | <1 | <1 | -- | -- | <1 | -- | <1 | <1 |
| Hg | Total | 0.1 | 0.2 | 0.6 | <0.1 | <0.1 | 0.4 | -- | -- | 0.2 | -- | 0.2 | 0.2 |
| | Diss. | 0.2 | 0.3 | 0.4 | 0.2 | 0.6 | 0.1 | -- | -- | <0.1 | -- | <0.1 | 0.2 |

* Units are in micrograms per litre.

Table A3
Concentrations of Pesticides in the Water and Sediment
of the Whitewater River at Towanda, Kansas, 1977*

| Constituent | Date and Time, hr | | | | | | | | 19 May Sediment |
|--------------------|-------------------|----------------|----------------|----------------|----------------|----------------|-------|----------------|--------------------|
| | 21 Apr 1110 | 19 May 1300 | 21 May 1500 | 21 May 1900 | 23 May 1325 | 11 Aug 1610 | 2015 | 12 Aug 1010 | |
| PCN | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | -- |
| Aldrin | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.10 |
| Lindane | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.10 |
| Chlordane | <0.10 | <0.10 | <0.10 | <0.10 | <0.10 | <0.10 | <0.10 | <0.10 | <1.00 |
| DDD | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.10 |
| DDE | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.10 |
| DDT | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.10 |
| Dieldrin | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.10 |
| Endosulfan | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | -- |
| Endrin | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.10 |
| Ethion | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.10 |
| Toxaphene | <0.10 | <0.10 | <0.10 | <0.10 | <0.10 | <0.10 | <0.10 | <0.10 | <1.00 |
| Heptachlor | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.10 |
| Heptachlor Epoxide | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.10 |
| PCB | <0.10 | <0.10 | <0.10 | <0.10 | <0.10 | <0.10 | <0.10 | <0.10 | <1.00 |
| Malathion | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <1.00 |
| Parathion | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.10 |
| Diazinon | <0.01 | <0.01 | 0.12 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.10 |
| Methyl Parathion | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.10 |
| 2,4-D | 0.08 | 0.13 | 0.30 | 0.22 | 0.32 | 0.45 | 0.27 | <0.01 | <1.00 |
| 2,4,5-T | 0.03 | 0.04 | 0.11 | 0.05 | 0.11 | 0.49 | 0.24 | <0.01 | <1.00 |
| Silvex | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <1.00 |
| Trithion | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.10 |
| Methyltrithion | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | 0.01 | <0.01 | <0.01 | <0.10 |

* Units are in micrograms per litre.

Table A4
General Water-Quality Data for the Whitewater River at Towanda, Kansas, 1974*

| Parameter | Date and Time, hr | | | | | | | | | | | |
|--|-------------------|----------------|----------------|----------------|----------------|----------------|---------------|----------------|------|------|----------------|------|
| | 21 Apr 1110 | 19 May 1300 | 21 May 1500 | 22 May 1700 | 27 Jun 1115 | 21 Jul 1330 | 4 Aug 1340 | 11 Aug 1610 | 2015 | 1010 | 12 Aug 1230 | 1410 |
| Discharge, cfs | 15 | 10 | 1880 | 2260 | 1360 | 760 | 503 | 19 | 88 | 2790 | 3150 | 3310 |
| Turbidity, JTU | 20 | 30 | 600 | 700 | 480 | 700 | 460 | 65 | 30 | 80 | 1200 | 190 |
| Conductivity, μ mhos | 2120 | 1820 | 445 | 419 | 285 | 293 | 320 | 815 | 1570 | 1530 | 310 | 210 |
| Temperature, °C | 16.0 | 22.5 | 18.0 | 17.5 | 17.0 | 18.0 | 18.0 | 23.0 | 28.0 | 24.0 | 21.0 | 21.0 |
| Dissolved oxygen | 7.2 | 9.2 | 4.6 | 4.1 | 6.3 | 6.4 | 8.0 | 7.0 | 14.4 | 8.6 | 4.4 | 4.6 |
| Biochemical oxygen demand | 0.9 | 3.5 | 3.4 | 5.2 | -- | -- | -- | -- | 6.4 | 5.5 | 5.0 | 4.4 |
| Chemical oxygen demand | 39 | -- | 210 | 140 | 130 | 160 | 150 | 45 | 46 | 36 | 200 | -- |
| Alkalinity as CaCO_3 | 260 | 230 | 71 | 90 | 64 | 67 | 71 | 180 | 180 | 190 | 71 | 48 |
| pH | 7.7 | 7.9 | 7.2 | 7.4 | 7.4 | 7.4 | 7.3 | 7.9 | 8.3 | 7.8 | 7.5 | 7.4 |
| HCO_3^- | 270 | 280 | 86 | 110 | 78 | 82 | 86 | 220 | 220 | 230 | 86 | 4.4 |
| CO_2 | 8.6 | 5.6 | 8.7 | 7.0 | 5.0 | 5.2 | 6.9 | 4.4 | 1.8 | 5.8 | 4.4 | 3.7 |
| Ca + Mg hardness | 830 | -- | 160 | 150 | 100 | 110 | 120 | 350 | 560 | 540 | 110 | -- |
| Noncarbonate hardness | 560 | -- | 89 | 63 | 39 | 44 | 48 | 170 | 380 | 350 | 39 | -- |
| Total dissolved solids | 1410 | -- | 273 | 249 | 181 | 204 | 222 | 559 | 1010 | 1010 | 198 | -- |
| Dissolved Ca | 240 | -- | 46 | 45 | 30 | 33 | 35 | 100 | 150 | 150 | 32 | -- |
| " Mg | 55 | 41 | 11 | 10 | 6.8 | 6.9 | 7.6 | 24 | 44 | 41 | 7.1 | -- |
| " Na | 120 | 120 | 22 | 16 | 12 | 13 | 14 | 39 | 120 | 110 | 20 | -- |
| " K | 6.6 | 7.1 | 6.6 | 7.5 | 6.4 | 6.3 | 6.3 | 6.7 | 5.6 | 4.7 | 5.1 | 4.6 |
| " Cl | 230 | 210 | 39 | 25 | 16 | 15 | 18 | 72 | 250 | 230 | 34 | -- |
| " F | 0.5 | 0.4 | 0.4 | 0.3 | 0.4 | 0.4 | 0.4 | 0.3 | 0.3 | 0.3 | 0.4 | -- |
| " SO_4 | 480 | 330 | 71 | 76 | 42 | 46 | 49 | 150 | 270 | 290 | 40 | -- |
| " SiO_2 | 13 | 15 | 9.6 | 9.4 | 9.2 | 11 | 11 | 15 | 9.9 | 7.8 | 7.7 | 8.4 |
| Total N | 1.6 | 2.2 | 7.0 | -- | 7.9 | 6.2 | 6.0 | 2.3 | 1.3 | 1.9 | 5.4 | 5.1 |
| Total Kjeldahl Nitrogen | 1.2 | 1.9 | 5.2 | -- | 6.0 | 4.1 | 3.7 | 1.2 | 1.3 | 1.6 | 4.3 | 3.7 |
| $\text{NO}_2 + \text{NO}_3 - \text{N}$ | 0.38 | 0.31 | 1.8 | 1.7 | 1.9 | 2.1 | 2.3 | 1.1 | 0.01 | 0.25 | 1.1 | 1.4 |
| Total P | 0.13 | 0.35 | 1.0 | 0.31 | 0.99 | 0.87 | 0.79 | 0.28 | 0.15 | 0.19 | 1.4 | 0.75 |
| Total Organic Carbon | 7.3 | 7.1 | 42 | 43 | 39 | 34 | 31 | 8.8 | 10 | 9.7 | 63 | -- |

* Units are in micrograms per litre except where indicated otherwise.

Table A5

Microbiological Data for the Whitewater River at Towanda, Kansas, 1977*

| <u>Date</u> | <u>Fecal Coliforms</u> | <u>Immediate Coliforms</u> | <u>Fecal Streptococci</u> |
|-------------|------------------------|----------------------------|---------------------------|
| 21 Apr | 64 | 11,000 | 300 |
| 19 May | 60** | 750 | 220 |
| 21 May | 100,000 | 140,000 | 170,000 |
| 21 May | 88,000 | 110,000 | 290,000 |
| 21 May | 110,000 | 150,000 | 370,000** |
| 27 Jun | 1,900 | 4,100 | 2,000 |
| 21 Jul | 18,000 | 13,000 | 440 |
| 4 Aug | 2,300** | 14,000 | 12,000 |
| 11 Aug | 76,000 | 460,000** | 160,000 |
| 11 Aug | 40,000 | 300,000 | 150,000 |
| 12 Aug | 93,000 | 320,000 | 220,000 |

* Units are colonies per 100 millilitres.

** Nonideal count.

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United States. Waterways Experiment Station, Vicksburg, Miss.
Preliminary evaluation of the water quality of proposed Towanda Lake, Kansas. Vicksburg, Miss. : Environmental Effects Laboratory, U. S. Army Engineer Waterways Experiment Station ; Springfield, Va. : available from National Technical Information Service, 1978.

28, 27, 6 p. : ill. ; 27 cm. (Miscellaneous paper - U. S. Army Engineer Waterways Experiment Station ; Y-78-3)

Prepared for U. S. Army Engineer District, Tulsa, Tulsa, Okla.

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TA7.W34m no.Y-78-3